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Greenhouse gases emission trade-offs for benefits gain - An analysis from paddy rice and upland crops cultivation in Hau Giang province, Viet Nam

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ABSTRACT

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Keywords

Black sesame, corn, greenhouse gases emission, mungbean, paddy rice Agriculture activities require energy for operation and emit greenhouse gases (GHGs) into the atmosphere. However, agriculture provides essential nutrients with carbon sources through its main and by-products. This study used the life cycle assessment methodology to evaluate the carbon balance in agricultural systems of paddy rice (PR), corn, mung bean (MB), and black sesame (BS) in the summer-autumn growing season in the Vietnamese Mekong Delta. The results showed that PR and upland crops produced a net carbon source of 1,280.9–18,915 kg-C ha⁻¹. Corn cultivation achieved the best value in carbon index analyses. To have one calorie from grain, selected crops must trade off 115.19–501.81 mg-CO₂e. This study concluded that four selected crop cultivations achieved carbon analysis benefits. However, corn is a suitable recommendation for adapting to the agricultural conversion from PR farming to better upland crop cultivation.

1. INTRODUCTION

Although agriculture played an essential role in Vietnam's development, it must be restructured to achieve sustainable growth, reduce greenhouse gases (GHGs) emissions, avoid adverse environmental impacts, and utilize ecological benefits. In terms of environmental orientation, the agricultural restructure must focus on enhancing solutions for reducing GHG emissions.

Rice farming is a crucial sector of Vietnam's agriculture, with the Mekong Delta (MD) being the largest and most productive region for paddy rice (PR) cultivation in the country. However, rice production harms the environment by emitting GHGs. In 2021, rice production in Vietnam emitted 35.68 Mt-CO₂e (FAOSTAST database). By the year

2030, emissions from agriculture could reach 85 Mt-CO₂e per year (Hoa et al., 2014).

Enhancing the PR land system is essential for restructuring the crop industry and contributing to the successful implementation of increasing added value for agricultural products and sustainable agricultural development. In the period 2016–2020, about 510,000 ha of PR land was planned for conversion to upland crops or aquaculture combinations. In particular, within the MD, 84,000 ha of summer-autumn (S-A) PR farming area was scheduled for cultivating corn, sesame, vegetables, and other crops (Ministry of Agriculture and Rural Development of Vietnam, 2014).

In Hau Giang province, the government encourages the farmers to convert triple PR farming to other

agricultural models, including double PR-one cash crop farming. The total cultivated vegetable area increased by 119.0% in 2020 compared with 2019, in which the grown area of corn increased by 118.7% (Department of Agriculture and Rural Development of Hau Giang Province, 2020). In 2021, corn was cultivated over 2,930 ha and vegetables over 25,159 ha (Hau Giang Statistical Office, 2021).

It is suggested that PR monocultures harm the environment in the long term, while rotation could bring several benefits. Rowing upland crops such as corn or mung bean (MB) helps improve soil properties and plant growth (Linh et al., 2017). A study on GHGs emissions showed a reduction of 52–58% in total emission reduction by rotation with upland vegetables or lotus compared to triple PR farming (Dang et al., 2021).

Research on agriculture's carbon footprint has gained widespread attention, becoming a key focus in efforts to achieve sustainability and adapt to climate change. The carbon footprints of agriculture are measured as agricultural activities' impact on the environment in the number of inputs and outputs measured in the equivalent of CO₂. The optimal and efficient carbon usage in agriculture production will help succeed in the sustainable development goals (Ensure sustainable (SDGs) under SDG12 consumption and production patterns) through the sustainable reuse of agricultural residues. It could help achieve SDG13 (Climate action) by mitigating GHGs emissions to prevent climate change and global warming (United Nations, 2014).

In the period 1961–2014, GHG emissions from rice cultivation in Vietnam increased by 66% (Maraseni et al., 2018). The research on GHG emissions from upland crops in Vietnam is limited. Hung et al. (2021) reported that the emission from leafy vegetable cultivation, including choy sum

(132.6–215.6 kg-CO₂e t⁻¹), mustard (37.0 kg-CO₂e t⁻¹), and cabbage (91.4 kg-CO₂e t⁻¹). Liem and Phuoc (2023) showed the emission from sweet potato (15.74 t-CO₂e ha⁻¹ year⁻¹ and 0.26 kg-CO₂e t⁻¹) and taro (19.66 t-CO₂e ha⁻¹ year⁻¹ and 0.53 kg-CO₂e t⁻¹) cultivation in Kien Giang province.

This study used the life cycle assessment (LCA) methodology to compare PR and upland crop cultivation in the S-A growing season on carbon footprint, carbon efficiency, and carbon sustainability index. Besides that, the results will show the carbon trade-off for nutrient gain. They provide background information for the carbon footprint calculation of crop variety, supporting planning and decision-making for the agricultural sustainability transition.

2. MATERIALS AND METHOD

The pilot field experiments on corn, MB, and black sesame (BS) were separately conducted in Chau Thanh A district and Vi Thanh City, Hau Giang province, from March to June 2022. The total experiment area was 3,950 m² of corn, 1,800 m² of BS, and 2,350 m² of MB (Figure 1). Because we established only one model for each crop, we used the average values of agricultural inputs and outputs results from two experimented areas for analysis. Our experiments were conducted on Glevic Fluvisols, a specific soil classification (Hau Giang People Committee, 2022). To collect the PR cultivation data, we randomly selected and interviewed 240 households from February to March 2022 in Long My district, Hau Giang province. The surveyed sample size and the selection method were adapted to assess life cycle GHG emissions from the horticultural product, following the guide by the British Standards Institution (The British Standards Institution - BSI, 2012).



Figure 1. Research area and activities

The Walkley-Black method was used for plant biomass carbon analysis at the Laboratory of Environmental General Analysis, Center for Management of Practice and Analysis, Kien Giang University, Vietnam.

The LCA methodology was applied to estimate GHGs for producing and applying agricultural inputs. The system boundaries were set to 100 years, and all emissions were converted to CO_2 equivalents using the Intergovernmental Panel on Climate Change (IPCC) coefficients guideline (IPCC, 2013).

While using emission and conversion factors specific to Vietnam would be the most accurate approach for estimating GHG emissions and carbon sequestration, publications in this field from Vietnam remain limited. To our knowledge, many previous studies conducted in Vietnam used the emission factors/conversion factors from foreign countries (Phong & Loi, 2014; Le & Ha, 2015; Truong et al., 2017; Liem et al., 2022). We also used the default emissions from (IPCC, 2006) and published papers in this research.

The mass of by-products was estimated based on the crop-to-residues ratio (CRR). This study qualified straw, stover, cob, and shell of rice, corn, bean, and sesame through their dry grain. The CRR is as follows CRR_{Rice straw} = 1.53 (Purohit, 2009), CRR_{Corn} stover = 2.5 (Soni et al., 2013), CRR_{Corn cob} = 0.15 (Honorato-Salazar & Sadhukhan, 2020), CRR_{MB} stover = 1.35 (Wang et al., 2013), CRR_{MB shell} = 0.323 (Soni et al., 2013), CRR_{BS stover} = 3.8 (Honorato-Salazar & Sadhukhan, 2020), and CRR_{BS shell} = 1.86 (Ali & Jan, 2014).

$$C_{\text{Net}}$$
 (kg-C ha⁻¹) = C_0 (kg-C ha⁻¹) – C_I (kg-C ha⁻¹)

In which:

C₁ is total carbon inputs (kg-C ha⁻¹) based on CO₂e emission and was determined by C₁= CO₂e emission $\times 0.27$ (12/44 as the mass of C and CO₂); CO is total carbon outputs (kg-C ha⁻¹; including product and by-products) and was determined CO = Σ (Yj \times %Cj) (Yj: yield of plant j part – grain, stalk/straw, and shell in kg-C ha⁻¹; %C: carbon content of plant j part is presented in Table 1).

Outputs	%C References	Outputs	%C References
1. Rice		3. MB	
1.1 Grain	43.8–46.6 This study (Fu et al., 2012; Biswas et al.,	3.1 Grain	42.1-44.8 This study
1.2 Straw	45.7–61.4 2017; Maguyon-Detras et al., 2020)	3.2 Stover	45.7-47.3 This study
2. Corn		3.3 Shell	45.1-46.5 This study
2.1 Grain	46.2-46.8 This study	4. BS	
2.2 Stover	43.65–43.8 (Shuangning et al., 2005; Ioannidou et al., 2009)	4.1 Grain	40.8-42.4 This study
2.3 Cob	42.1–43.77 (Ioannidou et al., 2009; Biswas et al., 2017)	4.2 Stover	44.5–52.43 (Ateş et al., 2006; Khairy et al., 2023)
		4.3 Shell	43.98–44.02 (Sellami et al., 2008)

Table 1. The carbon content of outputs

Emission from inputs production and application (internal gasoline combustion) was qualified by the MiLCA software application (Japan Environmental Management Association for Industry - JEMAI, 2014).

GHGs emission from inputs application = Σ [m_k (kg-input_k ha⁻¹) × EF_k (kg-CO₂e kg-input_k⁻¹)

GHGs emission from manage soil = $\Sigma [m_q (kg-gas_q ha^{-1}) \times ConF_q (kg-CO_2e kg-gas_q^{-1})$

 m_k is the mass of input k; EF_k is the emission factor of inputs k production and application; m_q is the mass of soil emission from gas q (CH₄ and N₂O); ConF_q is the conversion factor of gas q to CO₂e (1 kg-N₂O = 265 kg-CO₂e and 1 kg-CH₄ = 28 kg-CO₂e) (IPCC, 2013).

 $\begin{array}{l} m_{N2O} \; (kg\text{-}N_2O\;ha^{-1}) = m_{N\text{-}CF} \; (kg\text{-}N\text{-}CF\;ha^{-1}) \times EF_z \\ (kg\text{-}N_2O\text{-}N\;kg\text{-}N\text{-}CF^{-1}) \times 1.57 \end{array}$

 $m_{N\text{-}CF}$ is the amount of N-CF applied; EF_z is the emission factor applied for farming system type [0,93 (Yan et al., 2003) - 1% N₂O-N for the upland crop (IPCC, 2006) and 0.59 (0,3% kg-N, IPCC) – 0.768 (Vo et al., 2020) kg N₂O ha⁻¹ season⁻¹ for PR farming]; 1.57 is the value of 44/28 – the mass of N₂O and N.

 m_{CH4} (kg-CH₄ ha⁻¹) = D (days) × EF_{CH4} (kg-CH₄ ha⁻¹ day⁻¹)

D is PR farming days; EF_{CH4} is 1.167–2.78 kg-CH₄ ha⁻¹ day⁻¹ (Sandin, 2005; Le & Ha, 2015; Vo et al., 2018). The EFs from Vietnamese PR farming were priority applied.

Carbon efficiency: $CE = C_0 (kg-C ha^{-1}) / C_I (kg-C ha^{-1})$

Carbon sustainability index: CSI = $[C_0 (kg-C ha^{-1}) - C_I (kg-C ha^{-1})] / C_I (kg-C ha^{-1})$

 $Nu_{j-g}(kcal ha^{-1}) = m_{j-g}(t ha^{-1}) \times NuC_{j-g}(kcal t^{-1})$

 $m_{j\text{-g}}$ is the mass of crop j grain (t ha^{-1}); NuC_{j\text{-g}} is the nutrient conversion of crop j grain (kcal t^{-1}). In which, we use a nutrient conversion base on grain weight to calculate nutrient gain, including 3,600 \times 10³ kcal t-rice^{-1}, 3,470 \times 10³ kcal t-MB grain^{-1} (Namiki, 1995), 3,356 \times 10³ kcal t-corn grain^{-1} (Nuss & Tanumihardjo, 2010), and 5,590 \times 10³ kcal t-BS grain^{-1} (Elleuch et al., 2011). In the case of PR, rice was 80.4% rice grain (data in this study).

Emission trade-offs for nutrient gains (g-CO₂e kcal⁻¹) = Emission for 1 tonnage of product (g-CO₂e t^{-1}) / Nutrient in one tonnage of grain (kcal t^{-1})

GHGs emission for fertilizer elements gain (En-FG and Em-FG) index calculation

The Em-FG index was recommended to clearly explain the relationship between how much emission trade-off and fertilizer elements are applied for one hectare of growing area to produce one tonnage of the main product (Liem et al., 2022).

Em-FG of growing area: Em-FG (kg-CO₂e kgelements $^{-1}$ ha⁻¹) = GHGs emission (kg-CO₂e ha⁻¹) / [nitrogen element (kg-N ha⁻¹) + phosphate element (kg-P₂O₅ ha⁻¹) + potassium element (kg-K₂O ha⁻¹)]

Em-FG of grain: Em-FG (kg-CO₂e kg-elements⁻¹ t⁻¹) = GHGs-FG (kg-CO₂e kg-nutrients⁻¹ ha⁻¹) / Productivity (t ha⁻¹).

3. RESULTS AND DISCUSSION

3.1. Crop yields and resource consumption

Table 2 summarizes the inputs requirement and product achievement from agricultural cropping systems. Corn was the highest total biomass selected crop and was 45,698 kg ha⁻¹. The corn grain yield was 12,520 kg ha⁻¹, which was 2.65 times lower than by-product yield and was 27.4% of total biomass. Rice was the second largest biomass in the

cultivation area, with 13,425 kg ha⁻¹. Rice grain was 5,306 kg ha⁻¹, accounting for 39.5% of total biomass. The total biomass of MB was 4,704 kg ha⁻¹ including 37.4% grain (1,760 kg ha⁻¹) and 2,944 kg ha⁻¹ of by-products. Although BS grain only accounted for 15.0% of total biomass, 920 kg ha⁻¹, the by-product yield of BS was 5,207 kg ha⁻¹. The total biomass was 6,127 kg ha⁻¹ higher than the MB biomass yield.

Table 2. Input-Output from one hectare of crop cultivation

	PR	Corn	MB	BS
A. Inputs				
1. Human labor (h)	55	200	278	316
2. Agricultural machines (h)	95	57	120	209
3. Fossil fuels and electricity				
3.1 Gasoline (l)	0	9.5	7.5	0
3.2 Diesel (1)	252.2	662.5	468.3	321.4
3.3 Electricity (kwh)	2.34	0	7.5	0
4. Chemical fertilizers (kg)	224.6	533.3	204.2	157.8
4.1 Nitrogen (kg-N)	124.5	233.2	95	71.5
4.2 Phosphate (kg-P ₂ O ₅)	64.7	137.6	68.6	50.1
4.3 Potassium (kg- K_2O)	35.4	162.5	40.6	36.2
5. Agrochemicals (kg)	10.8	18	27.1	28.6
6. Seeds (kg)	120	13	25	4
6.1 PR	120	-	-	-
6.2 Corn	-	13	-	-
6.3 MB	-	-	25	-
6.4 BS	-	-	-	4
B. Outputs				
1. Total biomass of rice (kg)	13,425	-	-	-
1.1 Grain	5,306	-	-	-
1.2 Straw	8,118	-	-	-
2. Total biomass of corn (kg)	-	45,698	-	-
2.1 Grain	-	12,520	-	-
2.2 Stover	-	31,300	-	-
2.3 Cob	-	1,878	-	-
3. Total biomass of MB (kg)	-	-	4,704	-
3.1. Grain	-	-	1,760	-
3.2 Stover	-	-	2,376	-
3.3 Shell	-	-	568	-
4. Total biomass of BS (kg)	-	-	-	6,127
4.1 Grain	-	-	-	920
4.2 Stover	-	-	-	3,496
4.3 Shell	-	-	-	1,711

Crop cultivation requires labor for almost activities from seedling/drilling, irrigation, fertilization, agrochemicals application, and harvest. In this study, BS required the highest total working hours with 316 h ha⁻¹ and was 5.75 times higher than the lowest crop care hours of PR farming (55 h ha⁻¹). MB and corn were the second largest group of labor requirements with 278 and 200 h ha⁻¹, respectively. Fossil fuel was used for machine operation, water pump, and agrochemicals application. At the same time, electricity was only used for water pumps. The corn farmers primarily used gasoline-propelled handheld reaper machines to cut corn stalks. The diesel water pumps were the most popular

agricultural pumps in the local area. Corn required the most diesel and gasoline in cultivation activities, with 662.5 and 9.5 1 ha⁻¹, respectively. MB cultivation required three types of power 468.3 ldiesel ha⁻¹, 7.5 l-gasoline ha⁻¹, and 7.5 kwh ha⁻¹. BS only required diesel fuel with 321.4 1 ha⁻¹. Agrochemical applications included herbicides, fungicides, and pesticides. BS and MB required total agrochemicals of 28.6 and 27.1 kg ha⁻¹ higher than the two other crops. Corn cultivation used 18 kg ha⁻¹, while rice consumed the lowest agrochemicals quantity (10.8 kg ha⁻¹). Corn cultivation used total fertilizers 2.4–3.4 times higher than other crops. Nitrogen fertilizer was the most consumed fertilizer, followed by phosphate fertilizer and potassium fertilizer. Although corn cultivation used much fertilizer, it gained high-yield grain.

3.2. GHGs emission

Rice and corn growing areas emitted GHGs (7,514.2 and 5,263.9 kg-CO₂e ha⁻¹) higher than MB and BS cultivation (3,122.8 and 2,368.8 kg-CO₂e ha⁻¹) (Figure 2). However, based on the grain yield crops, the emission of one tonnage of BS and MB grain (2,574.8 and 1,774.3 kg-CO₂e t⁻¹) was higher than rice grain and corn grain (1,416.1 and 420.4 kg-CO₂e t⁻¹) (Figure 2).



Figure 2. GHGs emission from agricultural activities

Soil management was the most significant emission source of rice farming (73%), while it accounted for 12.1–17.8% of other crop cultivation. From rice farming, soil emissions also included high emissions from CH₄ under the flooded cultivating condition. Irrigation was the highest proportion of MB, corn, and BS cultivation (36.0%, 33.5%, and 25.7%). In detail, irrigation emissions included electricity/fossil fuel produced emissions and fuels combusted emissions by water pumps. Fertilizer production's emission was the second significant source of all selected crop cultivated (12.0–34.4%). Land preparation activities through fuels consumption for agricultural machines operation were estimated. Emissions from agrichemicals (herbicides, insecticides, and fungicides) used included manufacturing production and gasoline consumed for sprayers. The emission from land preparation and agrochemicals use was highest in BS cultivation (17.5% and 18.4%) and lowest in rice farming (5.3% and 2.2%). Harvest and residue collection emission was estimated through fuels consumed by the combine harvester, self-propelled round straw baler machine (rice farming), and selfpropel handheld reaper machine (corn cultivation). The lowest accounted for 0.6% in corn cultivation, and the highest accounted for 3.2% in PR farming.

Soil emission was the significant emission source from PR cultivation, while irrigation and fertilizers were two essential sources for mitigating. The flooded farming condition of PR emitted a high rate of methane. It has been found that by applying biochar alone or alternate wetting and drying water management will decrease methane emissions (Mohammadi et al., 2016; Sriphirom et al., 2020). Additionally, organic practice with green manure incorporation for rice farming in the early midseason drainage can mitigate GHGs emissions (Toma et al., 2021). The deep-band fertilizer practice enhances the efficiency of nitrogen fertilizer application for corn (Wu et al., 2021). Depending on the acidic soil condition of Hau Giang province, the application of nitrification inhibitors fertilizer was recommended for improving the efficient use and potential to increase the upland crop yield (Thapa et al., 2016). Besides, applying nitrification inhibitors fertilizer also reduces N2O emission compared with conventional nitrogen fertilizers (Halvorson et al., 2014; Thapa et al., 2016). Changing the irrigation method to drip irrigation and fertigation can reduce N2O emissions (Yao et al., 2019). To reduce soil emission, enhance N fertilizer efficiency, and optimize irrigation activity to decrease total C input to get a better CSI of selected crop cultivation systems.

3.3. Input-output carbon

We found that PR farming required the highest carbon input (2,049.3 kg-C ha⁻¹), followed by corn cultivation (1,435.6 kg-C ha⁻¹), MB cultivation (851.7 kg-C ha⁻¹), and BS cultivation (646.0 kg-C ha⁻¹). Corn cultivation produced 23,350.6 kg-C ha⁻¹ was 3.5 times higher than rice farming (6,745.8 kg-C ha⁻¹), 8.3 times higher than BS cultivation (2,829.8 kg-C ha⁻¹), and 10.9 times higher than MB cultivation (2,132.6 kg-C ha⁻¹). Crops cultivation produced significant carbon sources through their

biomass, including grain and by-products. The grain carbon source was highest in corn (5,821.8 kg-C ha⁻¹, 28.6% total biomass carbon), followed by rice farming (2,398.4 kg-C ha⁻¹, 32.5%), MB cultivation (767.4 kg-C ha⁻¹, 35.6%), and BS cultivation (382.7 kg-C ha⁻¹, 11.6%). By-products carbon source was higher than the marketable products carbon. Rice straw carbon was 4,347.4 kg-C ha⁻¹ and was 1.8 times higher than the rice grain. Corn cultivation produced 14,483.8 kg-C ha⁻¹ through corn byproducts and was 2.5 times higher than corn grain. Carbon from by-products of MB and BS cultivation were 1,365.2 and 2,447.1 kg-C ha⁻¹. They were 1.8 and 6.4 times higher than grain carbon sources (Table 2).

After combined harvesting, MB and BS seeds were separated, and around 2 cm or smaller biomass debris from stover and shells were sprayed on the topsoil. They were plowed into the soil to improve its fertility. Corn stover and rice straw are used for several purposes, including feedstock, mushroom cultivation, and soil orchard mulching. With any corn or rice cultivation area expansion, biomassproduced by-products will increase. This source will be a recycled material for other industry production, such as biomass gasification, biofuel, ethanol, and biochar.

3.4. Net carbon and CSI

The agroecosystem carbon analysis is presented in Table 3. The carbon footprint or net carbon, which was considered carbon sequestration, was highest in corn cultivation with 18,915 kg-C ha⁻¹ and was 4.0 times higher than PR farming (4,696.5 kg-C ha⁻¹), 8.6 times higher than BS cultivation (2,183.8 kg-C ha⁻¹), and 14.8 times higher than MB cultivation (1,280.9 kg-C ha⁻¹). Corn was the most carboneffective cultivated crop produced at 14.2 kg-C based on one kg-C input. The CE of rice and BS differed slightly with 3.3 and 4.4. MB cultivation was the minor CE crop that produced only 2.5 kg-C through one kg-C input. In this study, corn cultivation was the most sustainable selected crop, getting a CSI of 13.2. BS cultivation achieved the CSI at 3.4, higher than the 2.3 value of BS cultivation. MB cultivation was the lowest sustainable crop system in this study when CSI received 1.5.

Crop type					C_{I}		CNet	CE	CGI
		Low–High	High Medium Total		Low–High Medium		ha ⁻¹	CE	CSI
PR	Grain Straw	2,324.1–2,472.7 3,710.1–4,984.8	2,398.4 4,347.4	6,745.8	1,451.6– 2,647.1	2,049.3	4,696.5	3.3	2.3
Corn	Grain Stover Cob	5,784.2–5,859.4 13,662.5–13,709.4 790.6–822	5,821.8 13,678.1 805.7	20,350.6	1,426.3– 1,444.9	1,435.6	18,915	14.2	13.2
MB	Grain Stover Shell	741–788.5 1,085.8–1,123.8 256.4–264.3	767.4 1,104.8 260.4	2,132.6	847.9–855.4	851.7	1,280.9	2.5	1.5
BS	Grain Stover Shell	375.4–390.1 1,555.7–1,833 752.6 – 753.3	382.7 1,694.2 752.9	2,829.8	643.2–648.9	646.0	2,183.8	4.4	3.4

Table 3. Input-output carbon and carbon relationship of crop cultivation

3.5. Environmental trade-offs for grain yield and nutrient gain

In	S-A	growing	season,	corn	produce	d the	highest
nut	trients	through	corn gra	in at 4	457.0 ×	106 K	cal ha ⁻¹ .

Rice provided 149.7 × 106 Kcal ha⁻¹ and was 1.5 times higher than nutrients from MB grain (101.7 × 106 Kcal ha⁻¹). One hectare of BS cultivation supplied 51.4×106 Kcal (Figure 3).



Figure 3. Nutrients gain from crops grain

Producing one calorie from grain, corn cultivation only traded off 115.19 mg-CO₂e, while PR needed to trade off 501.81 mg-CO₂e. MB had to trade off higher GHGs than BS to produce one calorie (406.61 and 306.97 mg-CO₂e Kcal⁻¹) (Figure 4).

Rice and corn are the most cereal crops worldwide. In Asia, MB is a future innovative food crop (Li & Siddique, 2020). BS is an oil-rich oil vegetable and an animal feedstock supplied through its meal after extracting oil. Parallelly providing nutrients for world consumption, crop cultivation impacted the environment through several inputs used to produce and apply. Selected crop cultivation emitted a large amount of GHGs into the atmosphere through their production. This study showed the GHGs emission trade-off for crop nutrient gain. Although the previous studies compared data could not be found, this research result was the first step to improving the negative trade-off for positive gain. However, with a whole agricultural ecosystem of crop cultivation, it is possible to conclude that four selected crop cultivations had benefited in carbon net. Crop by-products provided a huge biomass source for secondary use - recycling activities that adapt to SDG12. In our study, we applied a range of emission rates followed by low, medium, and high intensity. The medium values were used for evaluation trade-offs and presented in Figure 4. The estimation of low and high GHGs emission exchange for benefits gain range from 1,003–1,829.2 kg-CO₂e t⁻¹ in PR farming, 417.7–423.2 kg-CO₂e t⁻¹ in corn cultivation, 1,766.4–1,782.2 kg-CO₂e t⁻¹ in MB production, and 2,563.5–2,586.1 kg-CO₂e t⁻¹ in planting BS. BS and MB cultivation has had a high range of trade-off for grain gain because of its low yield. While PR farming got a high yield, its total high emission also led this model to get a high index. High-value emissions from PR farming are primarily caused by flooded farming conditions and lead to methane gas emissions. In the case of nutrients, for achieving one calorie PR, corn, MB, and BS cultivation must swap 355.43-648.18, 114.44-115.93, 305.61-308.34, and 458.58-462.63 mg-CO₂e, respectively. Lower nutrient content in rice grain compared with other selected crop grains in this study causes the PR nutrient trade-off index to be higher than other crops. Corn production had a high yield and grain nutrient content, so this model cultivation consistently achieved a good trade-off index.



Figure 4. The environmental trade-off for grain yield and nutrient gain

4. CONCLUSION

PR is the most important crop in Vietnam and a staple in Mekong Delta agriculture. However, the practice of growing three rice crops annually has led to several negative environmental impacts. This study focuses on the Spring-Autumn growing season, aiming to recommend upland crop alternatives by examining carbon balance processes and nutrient outcomes. To assess the environmental impact of agricultural activities, the Life Cycle Assessment (LCA) methodology was applied. Overall, all crops studied showed positive results in achieving carbon targets. Corn is the best benefit upland crop that reaches the best value of carbon outcome, net carbon, CE, and CSI. One hectare of cultivated PR emitted the most jeopardous GHGs; however, through its high yield of grain and straw,

PR is the second-high benefit in the carbon balance, CE, and CSI.

To produce grain together with nutrient gain, corn must trade-off the lowest GHGs emission. This study is the first publication to show the GHG emissions trade-off index for nutrient achievement. Comparing PR and MB, one unit of rice weight is exchanged for a lower GHGs emission than MB. However, a higher GHGs emission traded off one PR calorie than MB.

People must trade off emissions to get nutrients from the grain. However, this study's results underline the benefit of produced carbon in the S-A agricultural systems case study in the MD. Additionally, to enhance the benefits of transitioning from rice production to other crops, corn is recommended as a viable alternative and could be strategically planned for cultivation

REFERENCES

Ali, S., & Jan, A. (2014). Sowing dates and nitrogen levels effect on yield attributes of sesame cultivars. *Sarhad Journal of Agriculture*, 30(2), 203–209.

Ateş, F., Pütün, A. E., & Pütün, E. (2006). Pyrolysis of two different biomass samples in a fixed-bed reactor combined with two different catalysts. *Fuel*, 85(12– 13), 1851–1859.

https://doi.org/10.1016/j.fuel.2006.01.015

Biswas, B., Pandey, N., Bisht, Y., Singh, R., Kumar, J., & Bhaskar, T. (2017). Pyrolysis of agricultural biomass residues: Comparative study of corn cob, wheat straw, rice straw and rice husk. *Bioresource Technology*, 237, 57–63. https://doi.org/10.1016/j.biortech.2017.02.046

Dang, K. K., Doan, M. T., Le, T. H. L., Nguyen, T. T. N., Pham, D. T., Do, H. T., Ngo, S. D., Vu, T. B. N., & Nguyen, P. A. (2021). Assessing the performance of climate smart rice production systems in the upper part of the Vietnamese Mekong River Delta. *Asian Journal of Agriculture and Development*, *18*(1), 15– 29. https://doi.org/10.37801/ajad2021.18.1.2

Department of Agriculture and Rural Development of Hau Giang Province. (2020). *Report on agriculture and rural development of Hau Giang province sector* - *The implementation in 2020 and mission direction in 2021.*

Elleuch, M., Bedigian, D., & Zitoun, A. (2011). Sesame (Sesamum indicum L.) seeds in food, nutrition, and health. In *Nuts and Seeds in Health and Disease Prevention* (pp. 1029–1036). https://doi.org/10.1016/B978-0-12-375688-6.10122-7

Fu, P., Hu, S., Xiang, J., Sun, L., Su, S., & Wang, J. (2012). Evaluation of the porous structure development of chars from pyrolysis of rice straw: Effects of pyrolysis temperature and heating rate. *Journal of Analytical and Applied Pyrolysis*, 98, 177–183. https://doi.org/10.1016/j.jaap.2012.08.005

Halvorson, A. D., Snyder, C. S., Blaylock, A. D., & Del Grosso, S. J. (2014). Enhanced-efficiency nitrogen fertilizers: Potential role in nitrous oxide emission mitigation. *Agronomy Journal*, 106(2), 715–722. https://doi.org/10.2134/agronj2013.0081

Hau Giang People Committee. (2022). Synthesis report: Hau Giang Province Planning for 2021 - 2030 and a Vision to 2050 (in Vietnamese).

Hau Giang Statistical Office. (2021). Report on the socio-economic situation of Hau Giang province in December, the fourth quarter and the year 2021. https://www.mpi.gov.vn/Pages/tinbai.aspx?idTin=53 289&idcm=503

Hoa, N. T., Hasegawa, T., & Matsuoka, Y. (2014). Climate change mitigation strategies in agriculture, forestry and other land use sectors in Vietnam. *Mitigation and Adaptation Strategies for Global* *Change*, *19*(1), 15–32. https://doi.org/10.1007/s11027-012-9424-0

Honorato-Salazar, J. A., & Sadhukhan, J. (2020). Annual biomass variation of agriculture crops and forestry residues, and seasonality of crop residues for energy production in Mexico. *Food and Bioproducts Processing*, *119*, 1–19. https://doi.org/10.1016/j.fbp.2019.10.005

Hung, N. P., Ampt, P., Rogers, G., & Ha, L. T. T. (2021). Preliminary N₂O emissions of major vegetable cropping systems in peri-urban Hanoi, Vietnam. *Vietnam Journal of Agricultural Sciences*, 4(4), 1257–1269. https://doi.org/10.31817/vjas.2021.4.4.05

Ioannidou, O., Zabaniotou, A., Antonakou, E. V., Papazisi, K. M., Lappas, A. A., & Athanassiou, C. (2009). Investigating the potential for energy, fuel, materials and chemicals production from corn residues (cobs and stalks) by non-catalytic and catalytic pyrolysis in two reactor configurations. *Renewable and Sustainable Energy Reviews*, 13(4), 750–762. https://doi.org/10.1016/j.rser.2008.01.004

IPCC. (2006). Guidelines for national greenhouse gas inventories (H. S. Eggleston, L. Buendia, K. Miwa, T. Ngara, & K. Tanabe (Eds.). The Institute for Global Environmental Strategies, Japan.

IPCC. (2013). Climate Change 2013: The physical science basis. Working group I contribution to the fifth assessment report of the Intergovernmental Panel on Climate Change (T. F. Stocker, D. Qin, G.-K. Plattner, M. Tignor, S. K. Allen, J. Boschung, A. Nauels, Y. Xia, V. Bex, & P. M. Midgley (Eds.). Cambridge University Press, Cambridge. https://doi.org/https://doi.org/10.1017/CBO9781107 415324

Japan Environmental Management Association for Industry - JEMAI. (2014). *The multiple interface life* cycle assessment (MiLCA) software (2.3). Toray Industries incorporated and Japan Environmental Management Association for Industry (JEMAI), Tokyo, Japan.

Khairy, M., Amer, M., Ibrahim, M., Ookawara, S., Sekiguchi, H., & Elwardany, A. (2023). The influence of torrefaction on the biochar characteristics produced from sesame stalks and bean husk. *Biomass Conversion and Biorefinery*, 0123456789. https://doi.org/10.1007/s13399-023-03822-9

Le, T. P., & Ha, M. T. (2015). Environmental effects of three rice production models of Small farmer(s)-Large field, GAP, and Conventional farming in the Mekong Delta. *Can Tho University Journal of Sciences*, 38(2), 64–75.

Li, X., & Siddique, K. H. M. (2020). Future smart food: Harnessing the potential of neglected and underutilized species for zero hunger. *Maternal and* *Child Nutrition*, *16*(S3), 1–22. https://doi.org/10.1111/mcn.13008

Liem, L. T. T., & Phuoc, N. T. K. (2023). Greenhouse gases emission from root vegetables cultivation in Viet Nam Mekong Delta: Cases study of sweet potato and taro in Kien Giang Province. In *Research Works by Japan Young Alumni (2022)* (pp. 83–91). VNU Publishing House.

Liem, L. T. T., Tashiro, Y., Tinh, P. V. T., & Sakai, K. (2022). Reduction in greenhouse gas emission from seedless lime cultivation using organic fertilizer in a province in Vietnam Mekong Delta Region. *Sustainability*, 14, 6102. https://doi.org/10.3390/su14106102

Linh, T. B., Guong, V. T., Tran, V. T. T., Van Khoa, L., Olk, D., & Cornelis, W. M. (2017). Effects of crop rotation on properties of a Vietnam clay soil under rice-based cropping systems in small-scale farmers' fields. *Soil Research*, 55(2), 162–171. https://doi.org/10.1071/SR16123

Maguyon-Detras, M. C., Migo, M. V. P., Van Hung, N., & Gummert, M. (2020). Thermochemical conversion of rice straw. In M. Gummert, N. Van Hung, P. Chivenge, & B. Douthwaite (Eds.), *Sustainable Rice Straw Management* (pp. 43–64). Springer Nature Switzerland AG. https://doi.org/10.1007/978-3-030-32373-8_4

Maraseni, T. N., Deo, R. C., Qu, J., Gentle, P., & Neupane, P. R. (2018). An international comparison of rice consumption behaviours and greenhouse gas emissions from rice production. *Journal of Cleaner Production*, 172, 2288–2300. https://doi.org/10.1016/j.jclepro.2017.11.182

Ministry of Agriculture and Rural Development of Vietnam. (2014). Decision on "Approving the planning for plants structure conversion based on paddy rice land for the period of 2014–2020" (No. 3367/QĐ-BNN-TT).

Mohammadi, A., Cowie, A., Anh Mai, T. L., De La Rosa, R. A., Kristiansen, P., Brandão, M., & Joseph, S. (2016). Biochar use for climate-change mitigation in rice cropping systems. *Journal of Cleaner Production*, *116*, 61–70.
https://doi.org/10.1016/j.jolanae.2015.12.092

https://doi.org/10.1016/j.jclepro.2015.12.083

Namiki, M. (1995). The chemistry and physiological functions of sesame. *Food Reviews International*, 11(2), 281–329. https://doi.org/10.1007/BF02640380

Nuss, E. T., & Tanumihardjo, S. A. (2010). Maize: A paramount staple crop in the context of global nutrition. *Comprehensive Reviews in Food Science* and Food Safety, 9(4), 417–436. https://doi.org/10.1111/j.1541-4337.2010.00117.x

Phong, L. T., & Loi, P. T. (2014). Environmental impact assessment of pummelo (Citrus maxima Merr.) and mango (Mangifera indica L.) production in the Mekong Delta. *Can Tho University Journal of Science*, 31, 39–50. Purohit, P. (2009). Economic potential of biomass gasification projects under clean development mechanism in India. *Journal of Cleaner Production*, 17(2), 181–193. https://doi.org/10.1016/j.jclepro.2008.04.004

Sandin, S. (2005). Present and future methane emission from rice fields in Dong Ngac commune, Hanoi, Vietnam. In *Earth Sciences Center*, *Göteborg University*. Göteborg University.

Sellami, F., Jarboui, R., Hachicha, S., Medhioub, K., & Ammar, E. (2008). Co-composting of oil exhausted olive-cake, poultry manure and industrial residues of agro-food activity for soil amendment. *Bioresource Technology*, 99(5), 1177–1188. https://doi.org/10.1016/j.biortech.2007.02.018

Shuangning, X., Weiming, Y., & Li, B. (2005). Flash pyrolysis of agricultural residues using a plasma heated laminar entrained flow reactor. *Biomass and Bioenergy*, 29(2), 135–141. https://doi.org/10.1016/j.biombioe.2005.03.002

Soni, P., Taewichit, C., & Salokhe, V. M. (2013). Energy consumption and CO₂ emissions in rainfed agricultural production systems of Northeast Thailand. *Agricultural Systems*, *116*, 25–36. https://doi.org/10.1016/j.agsy.2012.12.006

Sriphirom, P., Chidthaisong, A., Yagi, K., Tripetchkul, S., & Towprayoon, S. (2020). Evaluation of biochar applications combined with alternate wetting and drying (AWD) water management in rice field as a methane mitigation option for farmers' adoption. *Soil Science and Plant Nutrition*, 66(1), 235–246. https://doi.org/10.1080/00380768.2019.1706431

Thapa, R., Chatterjee, A., Awale, R., McGranahan, D. A., & Daigh, A. (2016). Effect of Enhanced efficiency fertilizers on nitrous oxide emissions and crop yields: A meta-analysis. *Soil Science Society of America Journal*, 80(5), 1121–1134. https://doi.org/10.2136/sssaj2016.06.0179

The British Standards Institution - BSI. (2012). Assessment of life cycle greenhouse gas emissions from horticultural products: Supplementary requirements for the cradle to gate stages of GHG assessments of horticultural products undertaken in accordance with PAS 2050 (PAS 2050-1). BSI Standards Limited.

Toma, Y., Takechi, Y., Inoue, A., Nakaya, N., Hosoya, K., Yamashita, Y., Adachi, M., Kono, T., & Hideto, U. (2021). Early mid-season drainage can mitigate greenhouse gas emission from organic rice farming with green manure application. *Soil Science and Plant Nutrition*, 67(4), 482–492. https://doi.org/10.1080/00380768.2021.1927832

Truong, T. T. A., Fry, J., Van Hoang, P., & Ha, H. H. (2017). Comparative energy and economic analyses of conventional and System of Rice Intensification (SRI) methods of rice production in Thai Nguyen Province, Vietnam. *Paddy and Water Environment*, 15(4), 931–941. https://doi.org/10.1007/s10333-017-0603-1

- United Nations. (2014). The road to dignity by 2030: ending poverty, transforming all lives and protecting the planet - Synthesis report of the Secretary-General on the post-2015 agenda. https://www.un.org/disabilities/documents/reports/S G_Synthesis_Report_Road_to_Dignity_by_2030.pdf
- Vo, T. B. T., Wassmann, R., Tirol-Padre, A., Cao, V. P., MacDonald, B., Espaldon, M. V. O., & Sander, B. O. (2018). Methane emission from rice cultivation in different agro-ecological zones of the Mekong River Delta: seasonal patterns and emission factors for baseline water management. *Soil Science and Plant Nutrition*, 64(1), 47–58.

https://doi.org/10.1080/00380768.2017.1413926

Wang, X., Yang, L., Steinberger, Y., Liu, Z., Liao, S., & Xie, G. (2013). Field crop residue estimate and availability for biofuel production in China. *Renewable and Sustainable Energy Reviews*, 27(2), 864–875. https://doi.org/10.1016/j.rser.2013.07.005

- Wu, P., Chen, G., Liu, F., Cai, T., Zhang, P., & Jia, Z. (2021). How does deep-band fertilizer placement reduce N₂O emissions and increase maize yields? *Agriculture, Ecosystems and Environment*, *322*(June), 107672. https://doi.org/10.1016/j.agee.2021.107672
- Yao, Z., Yan, G., Wang, R., Zheng, X., Liu, C., & Butterbach-Bahl, K. (2019). Drip irrigation or reduced N-fertilizer rate can mitigate the high annual N₂O+NO fluxes from Chinese intensive greenhouse vegetable systems. *Atmospheric Environment*, 212(February), 183–193. https://doi.org/10.1016/j.atmosenv.2019.05.056

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